Cyanide hydrolysis in a cyanide-degrading bacterium, \textit{Pseudomonas stutzeri} AK61, by cyanidase

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The cyanide-degrading bacterial strain AK61 was isolated from waste water at a metal-plating plant. The isolated strain was characterized by Gram-staining, quinone analysis, fatty acid profile and the API ZONE identification system, and identified as \textit{Pseudomonas stutzeri}. Whole cells were able to degrade cyanide rapidly in a 1 mM solution containing no organic substances, and produced ammonia as a product. The induction of the cyanide-degrading activity of \textit{P. stutzeri} AK61 did not depend on the presence of cyanide in the culture medium during growth. The cyanide-degrading enzyme was purified approximately 49-fold from a cell extract of \textit{P. stutzeri} AK61. The enzyme had a $K_m$ of 1.7 mM for cyanide and a specific activity of 54.6 nmol ammonia produced min$^{-1}$. The activity of the enzyme was optimal at 30°C and pH 7.5. The results of SDS-PAGE, gel-filtration chromatography and NH$_2$-terminal amino acid sequence analysis of the enzyme indicated that the functional enzyme was an aggregated protein consisting of a 38 kDa polypeptide. Like cyanidase (cyanide dihydratase), it was shown that the enzyme catalysed the hydrolysis of cyanide to ammonia and formate.

Keywords: cyanide degradation, hydrolysis, cyanidase

INTRODUCTION

Cyanide is highly toxic to living organisms because it inactivates the respiration system by tightly binding to cytochrome c oxidase (Solomonson & Spehar, 1981). Despite the toxicity of cyanide, large amounts of it are used in the metal-plating, pharmaceutical and agricultural-chemical industries. Waste water containing cyanide from these industries must be treated so that it contains a low level (< 1 mg l$^{-1}$) of cyanide before being released into the environment (Raybuck, 1992). Currently, cyanide in waste water is removed by chemical methods such as alkaline chlorination, ozonation and wet-air oxidation (Raybuck, 1992). These methods are expensive and require the use of hazardous reagents such as chlorine and sodium hypochlorite. Biological treatment of cyanide may be cheaper and more environmentally acceptable than chemical methods (Dubey & Holmes, 1995).

To develop the biological treatment of cyanide, screening of cyanide-degrading micro-organisms has been carried out. Some biological pathways have been suggested for the degradation or detoxification of cyanide in bacteria and fungi (Knowles & Bunch, 1986; Raybuck, 1992; Dubey & Holmes, 1995). \textit{Pseudomonas fluorescens} NCIMB 11764 has been shown to grow in a medium containing cyanide as the sole nitrogen source (Harris & Knowles, 1983). \textit{P. fluorescens} NCIMB 11764 converts cyanide to ammonia and carbon dioxide. This reaction may proceed by means of NAD(P)H-dependent oxygenase and cyanase (Dorr & Knowles, 1989). A \textit{Pseudomonas} species which degraded cyanide by another pathway was isolated from the activated sludge at a coke plant (White et al., 1988). Although the cyanide-degrading enzyme was not purified, it was shown that the strain degraded cyanide to ammonia and formate under both aerobic and anaerobic conditions. Recently, \textit{Alcaligenes xylosoxidans} subsp. \textit{denitrificans} DF3 (Ingvorsen et al., 1991) and \textit{Bacillus pumilus} C1 (Meyers et al., 1991) were also shown to degrade cyanide by the same pathway as the pseudomonad. These strains have been reported to degrade cyanide by cyanidase (cyanide dihydratase), which catalyses the hydrolysis of cyanide to ammonia and formate (Ingvorsen et al., 1991; Meyers et al., 1993). Some phytopathogenic fungi,
including *Gloeocercospora sorghi* (Fry & Munch, 1975; Wang et al., 1992), *Fusarium lateritium* (Cluness et al., 1993) and *Stemphylium loti* (Fry & Millar, 1972), have also been reported to degrade cyanide. Cyanide hydratases (EC 4.2.1.66) catalysing the hydrolysis of cyanide to formamide have been purified from *G. sorghi* and *F. lateritium*, and subsequently characterized.

In this study, we report the isolation of a cyanide-degrading bacterium from the waste water of a metal-plating plant. In addition, purification and characterization of a cyanide-degrading enzyme from the strain are described.

**METHODS**

**Isolation of the cyanide-degrading strain.** The cyanide-degrading micro-organism was isolated from the waste water of a metal-plating plant by cultivation in an isolation medium containing 1 mM KCN. The isolation medium contained the following ingredients per litre distilled water: 1 g KHPO₄, 2H₂O, 0.2 g MgSO₄, 7H₂O, 0.01 g CaCl₂, 0.01 g NaCl, 0.2 mg MnSO₄, 0.2 mg CuSO₄·5H₂O, 0.2 mg ZnSO₄, 2 g glucose, 1 g tryptone. The pH of the medium was adjusted to 7.6 using HCl. KCN solution, which was sterilized using a 0.2 μm pore filter, was added to the medium before inoculation. The isolation medium in a shake flask was inoculated with the waste water and incubated at 30 °C for 48 h with shaking. The grown culture was serially diluted and plated onto TYG agar (5 g tryptone 1⁻¹, 1 g yeast extract 1⁻¹, 1 g glucose 1⁻¹, 1 g KHPO₄ 1⁻¹, 20 g agar 1⁻¹). After 24 h incubation at 30 °C, the isolates obtained by restreaking colonies onto TYG agar were inoculated into isolation medium containing 1 mM KCN and incubated at 30 °C for 24 h. To determine cyanide-degrading activity of the isolates, the concentration of residual cyanide in each culture medium was assayed by the method of Lambert et al. (1975).

The cyanide-degrading bacterium was characterized by Gram-staining, quinone analysis, fatty acid profile and the API 20NE identification system (Bio Mérieux). Gram-staining was performed according to Gerhardt (1981). Fatty acids were analysed according to the method reported previously (Ikemoto et al., 1978a, b; Oyaizu & Komagata, 1981) and the quinone system was determined by the method of Yamada et al. (1969). The API 20NE identification system was inoculated and interpreted according to the manufacturer's instructions.

**Assay of cyanide-degrading activity of whole cells.** The isolated strain was grown in isolation medium with or without 1 mM KCN for 24 h and harvested in the exponential growth phase by centrifugation at 4000 g for 15 min. The pellet was washed twice in 50 mM phosphate buffer (pH 7.6) and resuspended in the same buffer. A suspension of washed cells was diluted in phosphate buffer to produce a solution with an OD₆₀₀ of 1 (DU 7400 spectrophotometer, Beckman). KCN as a substrate was added to a final concentration of 1 mM in the suspension and the mixture was incubated at 30 °C. Samples were removed from the reaction mixture at appropriate intervals and the reaction was stopped by the addition of 2 vol. 1 M NaOH. The samples were then centrifuged at 10000 g and concentrations of cyanide and ammonia in the supernatants were assayed colorimetrically using the methods of Lambert et al. (1975) and Fawcett & Scott (1960), respectively.

**Enzyme purification.** The isolated strain was grown using 3 litre medium. The cells were harvested and washed as described above. The washed cell suspension was disrupted by sonication for 10 min on ice and centrifuged at 10000 g for 15 min at 4 °C. The supernatant was used for purification of the cyanide-degrading enzyme as a cell extract. All purification steps were carried out at 4 °C. The cell extract was brought to 35% saturation with (NH₄)₂SO₄ by the addition of crystals. The supernatant was recovered by centrifugation and brought to 45% saturation by further addition of (NH₄)₂SO₄. The pellet was recovered and dissolved in 50 mM phosphate buffer (pH 7.6). This solution was applied to a DEAE column (Yamazen) previously equilibrated with 0.4 M NaCl in phosphate buffer. Proteins bound to the column were eluted with 0.5 M NaCl. An equal volume of 18 M (NH₄)₂SO₄ in phosphate buffer was added to eluted proteins, and the mixture applied to a phenyl-Toyopearl 650 column (Tosoh) equilibrated with 0.9 M (NH₄)₂SO₄ in phosphate buffer. Fractions eluted with decreasing (NH₄)₂SO₄ concentrations of 0.54-0 M were collected. Two volumes of 1 M phosphate buffer (pH 7.0) were added to the eluted proteins. These proteins were applied to a DEAE-Toyopearl 650 column (Tosoh) equilibrated with 0.15 M NaCl in 1 M phosphate buffer (pH 7.0). The proteins were eluted using a NaCl gradient from 0.15 to 1 M. All active fractions were applied to a Macro-Prep ceramic hydroxyapatite column (Bio-Rad) equilibrated with 0.3 M NaCl in 1 M phosphate buffer (pH 7.0) and eluted with a phosphate gradient of 1-250 mM. The active fractions were referred to as the purified enzyme. For gel-filtration chromatography, the purified enzyme was applied to a TSK-GEL G2000SWXL column (Tosoh), which had an exclusion limit of 100 kDa, and eluted with 50 mM phosphate buffer (pH 7.6) containing 0.3 M NaCl.

**Assay of enzyme activity.** Enzymic activity was assayed in 1.4 ml 50 mM phosphate buffer (pH 7.5) containing 10 mM KCl as the substrate. The reaction was carried out at 30 °C for 3-10 min. The amount of ammonia produced in the reaction mixtures was estimated using the method mentioned above. The amounts of protein were estimated using the method of Bradford (1976) with bovine serum albumin as the standard. One unit of the enzyme is defined as the amount needed to catalyse the formation of 1 μmol ammonia from cyanide min⁻¹ under the above conditions.

**Other assays.** The production of formate via the degradation of cyanide by the purified enzyme was assayed. The reaction was carried out in 50 mM phosphate buffer (pH 7.6) with an initial concentration of 0.1-1 mM cyanide at 30 °C. After degradation of cyanide in the reaction mixtures, concentrations of formate were determined. Formate was colorimetrically estimated by the reduction of NAD⁺ with formate dehydrogenase (Boehringer Mannheim) as described by Quayle (1966).

**SDS-PAGE.** SDS-PAGE was performed on 12% polyacrylamide gels according to the method of Laemmli (1970). Proteins were stained with 0.2% Coomassie brilliant blue R-250 in 50% ethanol/10% acetic acid. The molecular mass of the enzyme polypeptide was determined by comparison with the mobilities of standard proteins.

**RESULTS AND DISCUSSION**

**Isolation and characteristics of *P. stutzeri* AK61**

Several bacterial strains were isolated from the waste water of a metal-plating plant with the isolation medium containing 1 mM KCN. One of them, designated AK61, exhibited cyanide-degrading activity. Strain AK61 was a...
A degradation profile of cyanide by the washed cells was assayed in phosphate buffer containing 1 mM KCN (Fig. 2). Rapid removal of cyanide and the simultaneous accumulation of ammonia was observed in the reaction mixtures (data not shown). The final molar yield of ammonia was approximately 73%. Some bacterial strains which are able to degrade cyanide and generate ammonia have been isolated. P. fluorescens NCIMB 11746 converted cyanide to ammonia and carbon dioxide (Dorr & Knowles, 1989). In addition, Pseudomonas sp. (White et al., 1988), A. xylosoxidans subsp. denitrificans DF3 (Ingvorsen et al., 1991) and B. pumilus C1 (Meyers et al., 1991) were reported to convert cyanide to ammonia and formate. The cyanide-degrading activity of P. fluorescens NCIMB 11746 and A. xylosoxidans subsp. denitrificans DF3 was induced by cyanide during growth. On the other hand, induction of the cyanide-degrading activity of B. pumilus C1 was inhibited by addition of cyanide into the medium (Meyers et al., 1991). The cyanide-degrading activity of B. pumilus C1 was observed in only late-exponential- or early-stationary-phase cells cultured without cyanide. In contrast, the cyanide-degrading activity of P. stutzeri AK61 was observed whether it was cultured in cyanide-containing media or not (Fig. 2). The degradation profile of cyanide by the cells cultured without cyanide was similar to that of the cells cultured with cyanide. These results suggest that the biosynthesis of the cyanide-degrading enzyme by P. stutzeri AK61 is independent of the presence of cyanide during culture. This property of P. stutzeri AK61 may be useful for applying this strain to biological treatment of cyanide and preparing the cyanide-degrading enzyme, because P. stutzeri AK61 does not require cyanide to produce the cyanide-degrading activity and the induction is not inhibited by cyanide.

Purification of cyanide-degrading enzyme

A cell extract of P. stutzeri AK61 prepared from 3 l of culture in isolation medium without cyanide showed an initial specific activity of 1-11 units (mg protein)⁻¹. The cyanide-degrading enzyme was purified by salting-out and column chromatography. The results of a typical
Table 1. Purification of cyanide-degrading enzyme from *P. stutzeri* AK61

Reactions were carried out in 1.4 ml 50 mM phosphate buffer (pH 7.5) containing 10 mM KCN for 5 min at 30 °C.

<table>
<thead>
<tr>
<th>Fraction</th>
<th>Total protein (mg)</th>
<th>Total activity (units)*</th>
<th>Specific activity (units/mg protein)</th>
<th>Yield (%)</th>
<th>Purification (fold)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Cell extract</td>
<td>163.5</td>
<td>182</td>
<td>1.11</td>
<td>100</td>
<td>1-0</td>
</tr>
<tr>
<td>DEAM</td>
<td>1.41</td>
<td>47.2</td>
<td>33.5</td>
<td>26.0</td>
<td>30-1</td>
</tr>
<tr>
<td>DEAE-Toyopearl 650</td>
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<td>34.5</td>
<td>44.2</td>
<td>18.9</td>
<td>39.7</td>
</tr>
<tr>
<td>Hydroxyapatite</td>
<td>0.43</td>
<td>23.3</td>
<td>54.6</td>
<td>12.8</td>
<td>49.0</td>
</tr>
</tbody>
</table>

*One unit of the enzyme was defined as the amount needed to catalyse the formation of 1 μmol ammonia from cyanide min⁻¹.

Fig. 3. SDS-PAGE of steps during the purification of the cyanide-degrading enzyme from *P. stutzeri* AK61. Lanes: 1 and 8, molecular mass standards (rabbit muscle phosphorylase b, BSA, hen egg-white ovalbumin, bovine carbonic anhydrase, soybean trypsin inhibitor and hen egg-white lysozyme); 2, cell extract; 3, salt-out; 4, DEAM eluate; 5, phenyl-Toyopearl eluate; 6, DEAE-Toyopearl eluate; 7, hydroxyapatite eluate.

Fig. 4. NH₂-terminal amino acid sequence of the cyanide-degrading enzyme from *P. stutzeri* AK61 (AK61) and comparison with those of cyanidase from *A. xylosoxidans* subsp. *denitrificans* DF3 (DF3) and cyanide dihydratase from *B. pumilus* C1 (C1). Identical residues are enclosed by boxes.

Kinetic studies

The initial velocity of cyanide degradation by the purified enzyme was investigated over a substrate concentration range of 0.6–10 mM. A Lineweaver–Burk plot showed a linear response over this concentration range (data not shown). A Michaelis–Menten constant (Kₘ) of 1.7 mM was determined for cyanide. In order to study the effect of temperature and pH on the purified enzyme activity, it was assayed over a temperature range of 15–40 °C and a pH range of 6–10. The optimum temperature for the enzyme under our assay conditions was 30 °C (data not shown), above which the activity rapidly decreased. The optimum pH for the enzymic activity was 7.5 (data not shown), above which the activity decreased; no activity was observed at pH 10.

The effects on enzyme activity of the following metal ions were investigated: K⁺, Na⁺, Fe³⁺, Ca²⁺, Mg²⁺, Zn²⁺, Mn²⁺, Ni²⁺, Hg²⁺, Cr³⁺, Co²⁺. The activity was strongly inhibited by addition of Hg²⁺ at a concentration of 0.1 mM; the relative activity was 36%. This suggests the possible presence of thiol groups at the catalytic site of the enzyme. The addition of the other cations to the reaction buffer did not affect the activity of the enzyme.
Relative products

As a result of cyanide degradation by whole cells, ammonia was produced in the reaction mixture. The purified enzyme was able to degrade cyanide in a phosphate buffer without a co-enzyme. In addition, the NH$_3$-terminal amino acid sequence of the purified enzyme shows remarkable homology to that of cyanidase from A. xylosoxidans subsp. denitrificans (Ingversen et al., 1991). It is possible that the enzyme purified from P. stutzeri AK61 catalyses the same reaction as cyanidase (cyanide dihydratase), which catalyses the hydrolysis of cyanide to ammonia and formate (Ingversen et al., 1991; Meyers et al., 1993). To demonstrate that, like cyanidase, the purified enzyme converts cyanide to formate and ammonia, production of formate was assayed using formate-dehydrogenase-catalysed reduction of NAD$^+$

![Graph showing production of ammonia and formate after degradation of cyanide.](image)

**Fig. 5.** Production of ammonia and formate after degradation of cyanide. The reactions were carried out in 50 mM phosphate buffer (pH 7.5) with various concentrations of cyanide at 30°C for 30 min. After the cyanide was degraded, the concentrations of formate (○) and ammonia (■) in the reaction mixtures were assayed.

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REFERENCES


